

1 **Debunking trickle-down ecosystem services: the fallacy of omnipotent, homogeneous beneficiaries**

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3 **Abstract**

4 Ecosystem services research broadly assumes that an increased supply of nature’s goods and services
5 will yield increased benefits. We challenge this ‘trickle-down’ assumption by explicitly investigating the
6 factors that might impede ecosystem services yielding benefits to different stakeholder groups, based
7 on a targeted literature review of First Nations' access to shellfish on Canada's Pacific Coast. Our review
8 revealed four sets of barriers to realizing benefits from ecosystem services despite their abundance
9 within many First Nation territories. The barriers highlight problems of access, particularly as driven by
10 geographic location, technical capacity, markets and user conflicts, and management (of harvest and
11 access), all of which limit First Nations’ procuring of resources linked to key services. Our findings
12 demonstrate that simply increasing ecosystem service supply does not necessarily increase benefits for
13 individuals or groups. Realizing the promise that ecosystem services research will enhance human well-
14 being through improved management depends on the explicit consideration of how access mediates the
15 distribution of benefits.

16 **Keywords:** access; biophysical abundance; First Nations; food-social-ceremonial harvest; ecosystem-
17 based management; provisioning

18
19 **Introduction**

20 The concept of ecosystem services (ES) has become a valuable, widely used framework for
21 conceptualizing how people benefit from and depend on the diverse goods and services derived from
22 the biosphere (Lele et al. 2013). Much work has gone into mapping, quantifying and monetizing
23 ecosystem services and the benefits they yield. Such services have been categorized as provisioning
24 (e.g., provision of food and clean water); regulating (e.g., flood and disease risk mitigation); cultural
25 (e.g., contribution to spiritual, recreational, and cultural benefits); and supporting (e.g., nutrient cycling

26 and soil formation) (de Groot et al. 2002; Millennium Ecosystem Assessment 2005; Chan et al. 2012). In
27 other words, ecosystem services are the conditions and processes through which natural ecosystems
28 and their constituent species sustain and fulfill human life (Daily 1997).

29 A close reading of the ES literature reveals that the field as a whole effectively makes the
30 implicit assumption that increasing the supply of an ecosystem service (e.g., available shellfish) will
31 inevitably trickle down and lead to increased stakeholder benefits (e.g., via shellfish consumption). This
32 tacit assumption is manifest in the lion's share of ES literature addressing metrics of ES supply rather
33 than realized benefit (Tallis et al. 2012). Even those studies that assess the benefits or value of ES largely
34 ignore how those benefits might actually be realized, let alone how evenly they may be realized across
35 groups with diverse capabilities. While the importance of access to equitable ES benefit distribution was
36 articulated by the Millennium Ecosystem Assessment (2003), and the United Kingdom's Ecosystem
37 Services and Poverty Alleviation (ESPA) program provides a notable example where access is treated
38 explicitly, such insights about the true realization of benefits by diverse groups have not yet percolated
39 through the literature. An important component of how benefits are realized relates to the barriers that
40 impede diverse groups, and not just those less-developed nations, from realizing the benefits of ES. Such
41 barriers have yet to receive comprehensive treatment in the ES literature.

42 Despite this deficit, a broad diversity of research has provided a more nuanced perspective of
43 ecosystem service benefits. For example, several authors have pointed to the need to disaggregate ES
44 beneficiaries to articulate policy trade-offs, or identify winners and losers (Chan et al. 2007; Tallis et al.
45 2008; Daw et al. 2011; Ferraro & Hanauer 2011; Butler et al. 2013; Poppy et al. 2014). Others have
46 demonstrated the importance of space and scale to ES management, as potential beneficiaries ascribe
47 (cultural) value to and subsequently navigate space according to a particular social, economic, or
48 historical contexts (Alessa et al. 2008; Aswani & Lauer 2006; Dalton et al. 2010; Sherrouse et al. 2011;
49 Teh et al. 2012). ES benefits vary and are experienced differently depending on the scale of analysis

50 (Hein et al. 2006; Barbier et al. 2008; Carpenter et al. 2009; Martín-López et al. 2009; Plieninger et al.
51 2013), thus the need for ES trade-off and conflict analysis to be spatially and temporally explicit
52 (Douve & Ehler 2009; White et al. 2012). The linkages between ES and realized benefits, human well-
53 being, and poverty alleviation have also been explored (Brown et al. 2008; Daw et al. 2011; Polishchuk &
54 Rauschmayer 2012; Sikor 2013), with some pointing out that such benefits are inextricably linked to
55 intangible benefits or cultural ecosystem services which may require separate characterization (Chan et
56 al. 2011, 2012; Satterfield et al. 2013).

57 However, only a small number of papers drawing on the ES framework have explicitly accounted
58 for how access to ES and the associated benefits varies across space, groups, or communities (e.g.,
59 Martín-López et al. 2009; Daw et al. 2011; Hicks & Cinner 2014). The paucity of studies investigating how
60 benefits are realized by disadvantaged groups makes it clear that the field as a whole is still largely
61 based on the assumption that increasing ecosystem service supply will generally increase benefits. We
62 begin to address this gap in the literature using the case of Vancouver Island, British Columbia (BC), to
63 demonstrate the unavoidable link between the availability of an ES (in this case, shellfish for commercial
64 and subsistence harvesting), its distribution, and its promise of well-being, where well-being is linked in
65 part to the cultural benefits accrued as part of harvesting culturally valued food species.

66

67 **Erroneous Assumptions – the fallacy of ‘more is better’**

68 Shellfish harvesting continues to be important in maintaining cultural practices and livelihoods
69 of many First Nations groups along BC’s coast, and coastal Vancouver Island is no exception (Karpiak
70 2003; Ban et al. 2008; Menzies 2010). Culturally valued shellfish species on Vancouver Island include sea
71 urchins, chitons, snails, mussels, barnacles, abalone, geoduck, and a variety of clams. Accordingly, the
72 1990 Sparrow decision set out that “aboriginal rights to fish for food, social, and ceremonial purposes
73 have priority over all other uses of the fishery” (DFO 2008). A 2014 ruling by the Supreme Court upheld

74 this decision by asserting both the right to food fish and the right to fish for limited commercial activity
75 (The Globe and Mail 2014).

76 If the implicit assumption that an increased abundance of goods and services correlates directly
77 with increased benefits to communities holds true, then this court ruling could reasonably mean that
78 First Nations communities' ability to benefit from shellfish should now be closely linked to the
79 abundance of harvestable shellfish populations in the region. In other words, if benefits are not being
80 realized, then abundance must be the primary issue we address.

81 Acknowledging that low abundance of harvestable shellfish remains a significant issue on
82 northern Vancouver Island (Heaslip 2008), we nonetheless find that access is an equally large if not
83 larger problem. Drawing on the work of Ribot and Peluso (2003) and Brown et al. (2008) to characterize
84 *access*, we illustrate that many access barriers currently exist to limit the benefits First Nations
85 communities derive from this marine resource, *even where* shellfish abundance is high. We begin by
86 defining what we mean by access, and then define four sub-categories of barriers (geographical
87 location, technical capacity, markets and user conflicts, and management), each of which captures a
88 related set of circumstances that demonstrate the fallacy of trickle-down ecosystem services. This, in
89 turn, establishes the need to seriously consider the link between ES availability, access, and realized
90 benefits.

91

92 **Defining access (to ecosystem services)**

93 Access is listed by Schlager and Ostrom (1992) as one of five rights linked to resource control –
94 access, withdrawal, management, exclusion, and alienation. These are understood as the right to enter a
95 defined physical property and so the ability to withdrawal rights to obtain products or resources. For our
96 purposes, we use “access” to refer to spatial, legal, policy, or economic considerations that may prevent
97 individuals from realizing benefits from a supply of some good or service. Put simply, access is the “the

98 *ability to derive benefits from things*” (Ribot & Peluso 2003). In a coastal marine resource context,
99 Brown et al. (2008) further considered barriers to access that include permits and licenses, land
100 availability, gendered divisions of labour or gender-delimited access, geographical location, coastal
101 development, climate change, marine protected areas, technical capacity (fishing gear, boats, processing
102 facilities, storage equipment, skills), market isolation, user conflicts and pollution.

103

104 ***Geographic location: A resource does not benefit people if it is physically inaccessible.***

105 Two aspects of geographical location largely determine the degree of access and benefits
106 derived by coastal communities from marine resources. The first relates to physical and spatial barriers,
107 while the second relates to timing. Historically, restriction of physical proximity to coastal resources
108 began with the government instituted system of Indian Reserves along coastal BC that relocated and
109 restricted First Nations communities inland (Harris 2002; Heaslip 2008a; Joyce & Canessa 2009). This has
110 had numerous far-reaching effects, and its legacies continue to pose obstacles today.

111 There is much evidence that First Nations communities used to access marine resources by
112 locating villages nearby. Forced inland relocation means that many First Nations must now travel to
113 harvest resources that were once adjacent to their communities, or have completely lost access to areas
114 still considered part of their traditional territories. The problem is amplified by the fact that the
115 Department of Fisheries and Oceans Canada (DFO) will only allow communal commercial harvest for
116 “First Nations who have reserves fronting beaches with clam resources...” (DFO 2013). This problem
117 extends to many First Nations along BC’s coast, not just those on Vancouver Island, in particular those
118 whose reserves do not directly front beaches that contain harvestable shellfish or that define those
119 beaches as public, not reserve, space (Heaslip 2008a).

120 Such harvesting has always been seasonal because of winter weather and tides, and
121 contemporarily, some First Nations still maintain seasonal harvest camps, especially on the east coast of

122 Vancouver Island. But the reserve system and its accompanying economy (e.g., wage labour where
123 available) have made it increasingly difficult to accommodate temporal harvesting patterns based on
124 weather and seasonal abundance (Heaslip 2008a). Climate effects producing changing timing of
125 environmental conditions has only increased this post-colonial predicament (Turner & Clifton 2009).

126 Access on Vancouver Island is also undergoing ecological change as invertebrate populations,
127 including shellfish decline in the face of predation from recovering sea otter populations (Watson &
128 Estes 2011). Virtually no shellfish harvesting now occurs in some areas, although limited harvesting still
129 occurs mainly on small remote beaches by hand along the lower intertidal zone. Such areas must now
130 be accessed mainly by boat, except for the few clam beaches that are directly adjacent to reserves.
131 Consequently, other factors such as fuel or moorage costs have emerged with a disproportionately large
132 impact on the ability of people to harvest (Heaslip 2008a, Joyce & Canessa 2009).

133

134 ***Technical capacity: A resource does not benefit people if they lack the technical knowledge or capacity***
135 ***to harvest it.***

136 There are two aspects of technical capacity to access: the first relates to the technical
137 knowledge held by those who might benefit from a particular service, while the second considers
138 technical ability. Historically, digging for shellfish was a coordinated community effort involving a system
139 of hereditary chiefs, each head of a lineage or other kinship grouping with access rights to foreshore
140 areas. Gathering was done by hand, or with a trap set at low tide and then retrieved by foot or small
141 canoe at the next low tide (Menziés 2010). Many places along the BC coast were consistently used for
142 harvesting (McKechnie 2005), so much so that K'wakwaka'wakw groups created over 350 culturally
143 modified clam beaches and terraces (Harper et. al 1995). Indeed, recent research has shown that these
144 ancient clam gardens, a marine management tool used by First Nations groups all along the northwest

145 Coast of North America, actually increased the productivity of clams within them (Groesbeck et al.
146 2014).

147 Historical forces of colonialism thus cannot be over-stated, particularly as concerns their impact
148 on the food sovereignty and security of First Nations people throughout Canada and the world (Heaslip
149 2008a; Santha 2008; Parrish, Turner & Solberg 2007; Turner & Turner 2008). While, as noted above, the
150 reserve system resulted in a loss of immediate physical access to harvesting locations, the associated
151 residential school system removed many young people from their ancestral land and family (Turner &
152 Turner 2008). This in turn led to a loss of cultural knowledge important for the production, harvesting,
153 processing and use of food, as younger generations were unable to learn the specific technical
154 knowledge about what species, where, when, and how much to harvest (Turner & Clifton 2006; Parrish,
155 Turner & Solberg 2007). Where the transmission of such skills was broken, communities lost much of the
156 capacity to harvest resources in an appropriate fashion (Turner & Turner 2008), and so too a critical
157 cultural benefit of provisioning services. Additionally, current shellfish management boundaries drawn
158 by DFO do not respect First Nations' traditional territorial boundaries leading to a lack of coordination
159 and a break-down in resource governance protocols (Heaslip 2008a). One result is that where shellfish
160 harvesting does occur on Vancouver Island, it has become more spatially clustered:

161 *... many diggers describe a lack of communication and trust in the community. They argue that groups*
162 *heading out to dig clams generally do not know where other groups are planning to go. The*
163 *inexperienced end up heading to beaches where other groups' lights are visible. The result is increased*
164 *pressure on a beach that is already under use, while leaving other beaches without cultivation. There is a*
165 *need for coordination of the harvest, a role that, in the past, belonged to Chiefs (Heaslip 2008a).*
166

167 The second technical barrier to shellfish harvesting is the cost of the necessary technology to
168 reach harvesting areas and extract the resources efficiently. Harvesting now often requires a boat, fuel,
169 and maintenance costs at the very least (Castleden 2007). Expense prohibits many First Nations from
170 not only conducting subsistence harvesting but also from accessing many commercial fisheries. The
171 communal nature of shellfish harvesting on Vancouver Island, where groups of community members

172 often travel to clam beaches in one boat and work together to harvest the resource (Pinkerton and John
173 2008), mean that some *individual* costs and barriers to accessing the resource are reduced; the costs for
174 shellfish harvesting are therefore not as severe an access barrier as compared with some other types of
175 coastal BC fisheries.

176

177 **Markets and user conflicts: *A service does not benefit people if their harvest or sale is pre-empted***
178 ***or outcompeted by others.***

179 The actions of other resource users can impede a group's successful harvest or the market sale
180 of that harvest. Declines in coast-wide shellfish have been exacerbated by competition from commercial
181 and recreational harvesting. Conflicts between recreational harvesters on both the east and west coasts
182 of Vancouver Island have been identified (Joyce & Canessa 2009); as well as between First Nations
183 groups (Heiltsuk Tribal Council/DFO 2010/2011, DFO 2013). Many bands also note that areas 'dug out'
184 by commercial interests can no longer support subsistence harvest (Council of the Haida
185 Nation/Fisheries and Oceans Canada 2013).

186 The frustration over user conflict is exacerbated because entrance into the commercial
187 harvesting industry is expensive and complex, thus presenting a significant market barrier to First
188 Nations participation (Schreiber 2002; Heiltsuk Tribal Council/DFO 2010/2011; Fisheries and Oceans
189 Canada 2013). Many First Nations were and continue to be part of the commercial harvesting industry,
190 but declining stocks and more centralized license ownership means many First Nations have lost access
191 to commercial opportunities all along the BC coast (Schreiber 2002; Burke 2010). In BC, privatization of
192 fishing rights in the form of Individual Transferable Quotas (ITQs) that can be bought, sold, or traded,
193 now exist for most shellfish and many finfish fisheries (Ecotrust Canada 2004). These quotas are now
194 prohibitively expensive for most First Nations communities and rural residents (Ecotrust Canada 2004).
195 This ITQ system, with its high license and quota prices and government sponsored license retirement

196 and buyback programs, have caused many First Nations people to leave the fishery (Ecotrust Canada
197 2004). First Nations in this situation often face systemic disadvantages, where lower incomes and
198 property values and fewer economic opportunities overall leave them with less capital than many urban
199 residents. As of 2004, local residents on the west coast of Vancouver Island owned just two percent of
200 all ITQs in BC, while residents of Vancouver and Victoria owned 44 percent (Ecotrust Canada 2004).

201 Many First Nations communities on Vancouver Island have also identified commercial
202 aquaculture as a severe threat to wild harvesting (via spatial conflicts and diminished market
203 opportunities), though opinions vary considerably depending on community context (Joyce & Satterfield
204 2010). Shellfish and finfish aquaculture are rapidly growing industries in these regions. On the northwest
205 and northeast coasts of Vancouver Island, spatial analysis showed an increase of nearly 200% in beach
206 tenure area, representing a loss of 17% of total available harvestable area (Joyce & Canessa 2009). This
207 loss may well be exacerbated since such tenures are likely issued for some of the most highly
208 productivity areas.

209 Aquaculture often leads to the conversion of previously wild areas into privately leased tenures.
210 Processes of 'stealth privatization' have been identified as significant challenges facing shellfish
211 harvesters in the region and throughout Canada (Wiber et al. 2010). Private aquaculture leases typically
212 grant the lease-holder exclusive access to the land, water and foreshore (Deo 2002). This remains a
213 contentious issue on Vancouver Island, as treaties of lands and resources are still in negotiation.
214 Furthermore, private ownership and maintenance of lease space is expensive, limiting First Nations
215 access (Joyce & Canessa 2009).

216 User conflict can be considered a direct loss of access where these privately leased areas take
217 over wild clam and oyster bed areas. This occurs because leaseholders will generally choose to develop a
218 location that has proven biophysical capability (i.e., usually where FSC and wild clam harvest already
219 occurs) (Joyce & Canessa 2009). Indirect loss of access occurs where near-shore or deep-water shellfish

220 and finfish aquaculture infrastructure present hazards to navigation (Wiber et al. 2010). Finfish
221 aquaculture, in limiting First Nations people's access to digging beaches on the east coast of Vancouver
222 Island, for example, presents a direct barrier (Schreiber 2002; Joyce & Canessa 2009). Indirect loss of
223 access results from the perception that fish farms and associated waste are negatively impacting
224 surrounding clam beaches (Heaslip 2008b). The perception of contamination has been enough to
225 influence harvesters to avoid previous digging locations for 'fear of contamination' (Schreiber 2002;
226 Joyce & Canessa 2009).

227 Even after successful and legal capture of shellfish resources, benefits can be limited by the
228 access to markets for sale. Such market access can be impeded by lack of capital (credit and equipment),
229 geographic isolation (treated above), exclusionary practices, competition and conflict, state policy
230 support, licensing requirements and access fees (Ribot & Peluso 2003; Brown et al. 2008).

231

232 ***Management: The benefits of a service can be dramatically reduced if management structures***
233 ***for harvest and access are ineffective and under-resourced.***

234 Harvest rates, access, and catch limits are highly dependent on both the overarching structure
235 of management by DFO and the extent of communities' ownership of licenses in their area (Heaslip
236 2008a). Wild clam harvest for subsistence purposes is managed as a common property resource, with
237 limited licensing and seasonal openings and with the wild harvest licenses mostly owned by small-scale
238 users in rural communities (DFO 2013). In general, official limits in the form of daily quotas are attached
239 to communal licenses (although exact quantity restrictions vary between communities (DFO 2013)),
240 while mandated summer closures limit harvesting to the winter season (Beach 2002).

241 As of 2004, over 16% of coastal areas were officially closed to shellfish harvest in BC (Beach
242 2002). While these closures have often been necessary to protect public health, more recently, an out-
243 dated monitoring system (Beach 2002), and a lack of regional and local review of areas (Norman 2012)

244 have resulted in large areas (for instance Kyuquot Sound on the west coast of Vancouver Island)
245 remaining officially closed to harvest without timely reconsideration. Where First Nations communities
246 operate communal licences, communities must be able to afford continuous water quality assessments
247 in order to safely gather shellfish without fear of toxin contamination (Beach 2002; Heaslip 2008b;
248 Fisheries and Oceans Canada, 2013). If monitoring cannot be completed, areas remain closed until time
249 and budget constraints allow for re-sampling by the DFO.

250

251 **Discussion**

252 Our case study of shellfish harvesting by First Nations communities on coastal Vancouver Island,
253 Canada illustrates that increasing ecosystem service supply will not necessarily yield increased benefits
254 to communities, as much of the ES literature implies. We identified four classes of access barriers –
255 geographical location, technical capacity, markets and user conflicts, and management – that limit the
256 benefits First Nations communities do or potentially could realize from shellfish harvesting. Given the
257 shared colonial history and similar political and economic conditions today, these barriers are likely to
258 be relevant to North American west coast First Nations more broadly. Further, while it is important to
259 acknowledge that local context will influence the relevance of particular access barriers, the categories
260 of access barriers are not unique to BC Coastal First Nations (Hicks & Cinner 2014). A growing number of
261 studies are applying the categories identified by Ribot & Peluso (2003) to rural and resource-based
262 communities globally. For example, work by Hicks and Cinner (2014) identified the importance of key
263 access mechanisms (“rights-based, economic, knowledge, social, and institutional”) in mediating how
264 coral reef fishing communities in the Seychelles, Madagascar, Kenya and Tanzania are affected by the
265 distribution of ecosystem services. Using a similar framework, Milgroom, Giller and Leeuwis (2014) call
266 for critical work on the “dynamic relationships among quantity, quality and access” based on their
267 analysis of access barriers to natural resources faced by communities in southern Mozambique. These

268 studies contribute to a growing recognition that understanding the access barriers facing rural and
269 resource dependent communities is key to advancing our understanding of ecosystem service benefits
270 (Hicks and Cinner 2014).

271 The case of the Vancouver Island shellfish fishery offers a stark reminder of how management
272 decisions that fail to consider differences in beneficiaries' capacities to access resources can not only
273 limit the benefits derived from ecosystems but also fundamentally undermine the livelihoods of coastal
274 communities (Burke 2010). Crucially, it reminds us that ecosystem services yield not only quantitative
275 (i.e., a measure of abundance) but also qualitative (i.e., cultural and social dimensions) benefits and that
276 these interact to establish and maintain a sense of well-being. In general, we can point to five inter-
277 related constituents of well-being: security, basic material for a good life, health, good social relations
278 and freedoms and choice (Millennium Ecosystem Assessment 2005). A more nuanced perspective is
279 brought forward by the “capability approach” which expands on the "freedoms and choice" component
280 to argue that it is not just the end benefits that are important for well-being, but also the means by
281 which these are obtained (Abunge et al. 2013).

282 Subsistence harvesting provides both material and cultural benefits. For example, the
283 procurement of food provides an opportunity for the transmission of cultural knowledge, and other
284 cultural benefits accrue with the ability to feast and potlatch. These linked services or benefits yield
285 what are elsewhere referred to as tangible and intangible cultural ecosystem services, that is,
286 contributions from the ecosystem to non-material benefits that arise from human-ecosystem
287 relationships (Chan et al. 2012b). While difficult to value and highly subjective, they are crucial for the
288 health of First Nations communities. The presence of shellfish has been and continues to be culturally
289 important for First Nations communities; food and the processes of harvesting are intimately tied to
290 cultural practices, regeneration, and health (Karpiak 2003; Ban et al. 2008; Menzies 2010). Further, the
291 link between community sustainability and cultural practices (ceremony, storytelling, meaning-making,

292 identity, sense of place, relationship to nature, etc.) has been explored extensively (e.g., Barsh 1999;
293 Garibaldi & Turner 2004; Parrish, Turner & Solberg 2007; Turner & Turner 2008; Alfred 2009). Cultural
294 dislocation impacts First Nations at an individual and community level, limiting opportunities for self-
295 sufficient, healthy and autonomous lives. The conceptualization of barriers to subsistence harvesting, as
296 they pertain to benefits and more broadly, to well-being, must therefore include both the quantitative
297 (access to food) and the qualitative (access to the conditions conducive to realizing and maintaining
298 culture) aspects.

299 While the mechanisms by which the four factors pose barriers may be special to First Nations
300 communities, and to the BC coast, we would argue that all pertain to many ES in many contexts.
301 Geographic barriers—which are arguably the best recognized in the ES literature—can present
302 themselves much more subtly than relocation of villages, for example, by the shifting distribution of
303 ecological supply, or by a development (e.g., a road that separates a community from a site being
304 developed into a freeway). In each case, the service formerly available is effectively further away.
305 Technical and management barriers apply equally to regulating and supporting services: for example,
306 one cannot benefit from the provision of tidal energy for electricity production without the needed
307 know-how, financial capital, and appropriate infrastructure, or without the rights to install a tidal power
308 facility (or perhaps, sell electricity back to the grid). And even then, as is the case for some First Nations
309 considering this energy option, overcoming such technical barriers generally involves trading off
310 significant current and future cultural benefits (Rodman 2013). In short, because cultural services are so
311 frequently bundled with other ES (Klain & Chan 2012; Chan et al. 2012; Gould et al 2014), what applies
312 to provisioning services also applies to many cultural services.

313 In light of prevalent and interacting barriers to access, it is paramount for researchers who
314 frame their work in a context of human well-being to consider the cumulative impact the various
315 barriers have on the realization of benefits. Barriers often interact in a variety of ways to impede the

316 realization of benefits, sometimes synergistically. For instance, a *geographic barrier* (often the result of
317 colonization and the reserve system) that separates communities from their traditional fishing grounds
318 is exacerbated by a lack of employment and capital - *a technical barrier* - needed to navigate between
319 home and resource area. This is linked to a loss of traditional ecological knowledge and a gradual decline
320 in cultural practices, which feeds back into a reliance on non-traditional forms of subsistence practice
321 within a competitive market environment, often controlled by unfavourable management regimes
322 replete with *management, market and user conflict barriers*. An increase in provisioning goods (i.e.,
323 more shellfish) cannot therefore be assumed to trickle down to benefit all resource users equally.
324 Rather, for any particular resource context, there will likely be a specific set of barriers whose
325 interactions determine the contexts in which a given group will gain—or even lose (e.g., due to ‘elite
326 capture’)—from an increase in ES supply.

327 The importance of understanding access and non-monetary services and benefits is critical to
328 allow quantitative ES evaluation tools to inform understanding of trade-offs resulting from ecosystem
329 change (e.g., Integrated Valuation of Ecosystem Services and Tradeoffs (InVEST) - Guerry et al. 2012).
330 InVEST uses ecological production functions to estimate how changes in inputs (e.g., labour, materials,
331 habitat) affect service supply (e.g., fish). Social and economic indicators are then used to value outputs
332 (material and nonmaterial) and consider trade-offs from alternative uses of marine resources, *assuming*
333 *access*. While such tools represent a tremendous step forward in ES science, they currently consider
334 only a subset of barriers to access and therefore likely provide overly optimistic accounts of benefits to
335 some groups. Such tools may lend tacit support to the assumption that increased biophysical supply
336 necessarily generates more benefit to communities, giving too little attention to crucial institutional
337 factors mediating the realization of benefits.

338 Ultimately, a richer understanding of the benefits of ES requires jettisoning the assumption of an
339 omnipotent, homogenous beneficiary. The reality of heterogeneity in preferences and, more

340 importantly, ability to access and make use of ecosystem supply is often overlooked. If ES research
341 focuses on increasing supply without attending to the realities of realizing benefits, it may inadvertently
342 further privilege those (empowered) actors who are poised to benefit most from changes in supply,
343 potentially at the expense of disadvantaged parties. Explicitly considering the diversity and limitations of
344 potential beneficiaries will be instrumental in enabling the accurate assessment and mapping ES to yield
345 tangible and equitable improvements in ecological health and human well-being.

346

347 **Conclusion**

348 In the case of shellfish harvest by First Nations groups on the coast of Vancouver Island,
349 significant barriers to access include geographic location, technical capacity, markets and user conflicts,
350 and management. While increasing ecosystem service supply can be beneficial, our review shows that it
351 is unlikely to be sufficient to yield increased benefits to human coastal communities. A service does not
352 benefit people if they cannot physically access it; if they lack the technical knowledge or capacity to
353 harvest it; if they are pre-empted or outcompeted by others; or if over-arching management structures
354 are ineffective, under-resourced, or over-precautionary. For the ES literature to achieve its objectives of
355 improving management decisions and resulting human well-being, it must abandon the fallacy of trickle
356 down benefits, and begin to systematically address the diverse issues of access and how these differ
357 across individuals and groups.

358

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362

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